

## **DELIVERABLE 3.4**

# **MODEL INPUT OF SELECTED SCENARIOS**

Work Package 3 Measures & Pathways

30-06-2024











## Contents







## <span id="page-3-0"></span>1. LIST OF ABBREVIATIONS



*coverage (crop management)*







## <span id="page-4-0"></span>2. INTRODUCTION

## <span id="page-4-1"></span>1.1.The NAPSEA project

This project addresses the effectiveness of 'Nitrogen And Phosphorus load reduction measures from Source to sEA, considering the effects of climate change' (NAPSEA). The primary objectives of NAPSEA are to support national and local authorities in the selection of effective measures to reduce nutrient loads and to create political support for their implementation. The project applies an integrated approach spanning from pollution sources to sea, considering governance, nutrient pathways and measures, as well as ecosystem health. Geographically, the project focuses on the Wadden Sea catchment area, with specific case studies for the basins of rivers Rhine and Elbe, the catchment of the Hunze, as well as the Wadden Sea itself. NAPSEA serves as a platform to showcase practices in the implementation of socially acceptable, sustainable, and efficient measures.

The Work Package (WP) 3 aims to evaluate the connection between nutrient concentration and load reduction measures as well as the safe ecological boundaries for the Wadden Sea. The efficiency of nutrient reduction and enhanced retention measures will be assessed with a set of scenarios which has to be integrated into the modelling framework in order to prioritize mitigation measures under climate change.

## <span id="page-4-2"></span>1.2.Objectives

The assessment of the feasibility of measures (Gericke and Leujak 2023) concluded that scenarios should focus on agriculture, optimization of urban wastewater treatment plants (UWWTP), and nature-based solutions. The set of scenarios outlined by Gericke et al. (2024) comprise the wide range of specific measures [\(Table 1\)](#page-4-3). This report provides an overview of the input data used for the implementation of scenarios 1–4 into the modelling framework for rivers Elbe and Rhine (Task 3.5), and how the data was processed. Scenario 5 will be the combination of these scenarios. The "no measures" scenario 6 relies on climate scenarios and respective changes of the terrestrial water cycle already available at UFZ. The aim is to provide scenario data for the spatial units of the calibrated models (Musolff and Ledesma 2024) which includes the envisioned improvement of the mQM model in Germany. We included the Hunze catchment to allow for comparisons to its tailored, more detailed scenarios and datasets (cf[. Table 12](#page-23-1) in the Annex).

Depending on the scenario results and reduction needs to reach the safe ecological boundaries (cf. tasks 4.2 and 4.3), the use of additional data might be needed for scenario 7. Hence, we developed R scripts to process the different types of input data (vector/raster, numerical/categorical) as a flexible basis to integrate additional data or new insights at a later stage. These scripts also enable rapid updates if additional modelling units are delineated. Similar to the database for the calibrated and validated nutrient models (Musolff and Ledesma 2024), we selected European datasets which ideally cover the whole basins of rivers Elbe and Rhine, and national data only if unavailable.

Henceforth, we provide a concise overview of the current input and output files as well as of how the latter were created. Technical details like unit conversion (harmonization) or merging of multiple input files are neglected. We briefly discuss the steps to finalize the model input for the scenario runs and remaining issues.

<span id="page-4-3"></span>*Table 1. Proposed set of scenarios. All scenarios include climate change impacts in the river basins. More details in Deliverable D3.3 (Gericke et al. 2024).*







## <span id="page-5-0"></span>3. SCENARIO DATA

If not specified otherwise, the scenario data covers the current modelling units of mQM and the NUTS 3 regions for 2016 (Eurostat n.d.; 2018) which overlap with the reference catchment data for the WRRL (EEA 2020) and Switzerland (BAFU 2019) for the Rhine and Elbe basins as well as the Hunze catchment (Waterschap Hunze en Aa's 2016). The modelling units in the output files refer to the columns OBJECTID (Elbe, Rhine) and SiteID (Hunze). For the catchment of the Rhine gauge at Bimmen / Lobith, we set the OBJECTID to 6335060 (J. Ledesma, pers. comm.). The scenario[s 3.1](#page-5-1)[–3.5](#page-16-0) also include scenari[o 3.6.](#page-16-1)

### 3.1.Wastewater treatment

<span id="page-5-1"></span>Spatial coverage: DE x NL x Other x (no valid data for BE and FR)

Nutrient: N x P x

Scenario: Implementation of the provisional agreement on the new UWWTD reached by Council and Parliament (Council of the European Union 2024) in EU member states and the existing regulation in CH (Bundesrat 1999), both targeting at the retention and the concentration of total N and P in the outflow [\(Table 2\)](#page-5-2), no derogation and changes in the amount of treated wastewater assumed

<span id="page-5-2"></span>*Table 2. Target values implemented in the data processing. The current load was modified if the target was lower (outflow concentration) or higher (retention) than the current value. Of the two possible loads, the minimum was chosen if both targets apply, or the maximum otherwise.*



Data sources: Provisional UWWTD data reported under UWWTD data call 2021 (EEA 2023a), publicly available since January 2023, and a data table for Switzerland (P. Fischer, pers. comm.) based on the geodata model "Kläranlagendatenbank (ARA-DB)" (Federal Office for the Environment 2016)

Scenario data: Discharge points (longitude, latitude) with loads (t yr<sup>-1</sup>) for the reporting year (EU 2019-2020, CH 2020) and outflow loads for the unspecified scenario year, further attributes to filter (active/inactive discharge point, discharge to freshwater or other water-body types), separate files for EU member states and CH with Swiss ids of UWWTP and their discharge points being harmonized with EU scheme [\(Table 3\)](#page-6-1). The scenario results are exemplarily shown i[n Figure 1.](#page-6-2)

Data processing: For the spatial selection, the mQM modelling units were combined with the basins of Rhine and Elbe as well as the Hunze catchment. The discharge points (EU) and available coordinates (CH) falling into the two river basins were selected. For the EU data, the separate tables for discharge points, UWWTP, and agglomerations were joined. Agglomerations were grouped according to their generated load into the two size classes of the UWWTD (se[e Table 2\)](#page-5-2). As several agglomerations can belong to one discharge point, we determined the dominant size class based on the total generated load per size class. Based on this size class, the targets i[n Table 2](#page-5-2) were applied to the concentration and retention. For Switzerland, the design capacity was used to determine the size class. The retention was determined as 1 – outflow load / inflow load and the concentration as ratio of outflow load and treated wastewater. Any zero value was treated as missing value. In a few cases, implausible non-negative retention values and suspiciously high concentrations occurred but were neither removed nor altered.

NOTE: The original data does not contain valid load data for Belgium and France. Consequently, outflow loads are not (explicitly) considered in the model setup. Similarly, we have to assume unchanged outflow loads for the scenarios.





#### <span id="page-6-1"></span>*Table 3. Structure of output files for* wastewater treatment plants *with example values. Columns dcpState, dcpWB, and DateClosing should be used to exclude unsuitable data.*



\* Only EU data, \*\* based on generated load of agglomerations (EU) or design capacity (CH)



<span id="page-6-2"></span>*Figure 1. Scenario impact on outflow loads for the selected active UWWTP discharging into freshwater without Switzerland. The scenarios refer to the reference year ("current") as well as the UWWTD targets (concconcentration, ret-retention, target-overall target).*

## <span id="page-6-0"></span>3.2.Agricultural input

The agricultural input consists of the soil-surface N balances as the main source for N and the soil erosion for P. Accordingly, the first one is intended for the mQM model, the last one for the C<sup>n</sup>ANDY model. The scenarios [3.2.2](#page-9-0) and [3.2.3](#page-11-0) are mutually exclusive.

The agricultural P balances are not considered as the current annual balances are already close to zero, even negative. Furthermore, P accumulates in topsoil unlike N. The current values and the impact of planned measures are expected to be insufficient to (significantly) reduce the high P concentration in topsoil after decades of intensive fertilization (cf. Gericke and Leujak 2023).





### <span id="page-7-0"></span>3.2.1. Nitrogen balance

<span id="page-7-2"></span>Spatial coverage: DE x NL x Other

Nutrient: N x P

Scenario:

- DE: Anticipated trends in agriculture (Haß et al. 2022) and implementation of the current Fertilizer Ordinance (DüV 2021) until 2030
- NL: Anticipated trends in agriculture and implementation of the measures of the 7<sup>th</sup> Action Program until 2027 (cf. Table 6 in Gericke et al. (2023))

#### Data sources:

- DE: tabular N balances data at NUT 3 level provided by U. Häußermann (University Gießen, pers. comm.), calculated within the DüngEval project using the RegNBil approach (Häußermann et al. 2020)
- NL: modelled average nitrate leaching from agricultural land for different regions, sectors, and scenarios (van Boekel et al. 2021), the models LWKM, WOGWOD, and DSG were applied to derive the impact of the different measures on the nitrate concentration in leachate

Scenario data:

- DE: Area-weighted average of soil-balance N surplus including atmospheric deposition in kg/ha complemented by the atmospheric deposition and utilized agricultural area for modelling units. The processed data comprises three reference years (2019–2021) and the scenario year 2030 [\(Table 4\)](#page-8-0). The scenario effect strongly depends on the selected reference year given the recent decrease in N surplus [\(Figure 2\)](#page-7-1).

The values for the atmospheric deposition and utilized agricultural areas are separately provided to facilitate the harmonization with the data for scenari[o 3.3](#page-11-1) (for scenario [3.5\)](#page-16-0) and to adjust the N surplus for different agricultural land in the input for the calibrated mQM model and the scenario input.

NL: Average change in nitrate concentrations in leachate can be derived when mQM units for the Rhine basin are available. The relative change can either be used as a proxy for the change in N surplus or used to adjust the modelled nitrate concentration.



<span id="page-7-1"></span>*Figure 2. Distribution of average soil-surface N balance (surplus) for mQM modelling units in Germany, output for NAPSEA scenario 2 (column napsea\_scenario=2).*





<span id="page-8-0"></span>*Table 4. Structure of output files for the nitrogen balances with example values. Separate columns for N deposition and utilized agricultural area.*



Data processing:

- DE: The modelling units were intersected with the German NUTS 3 regions to obtain the area-weighted average values for the mQM model. The area-specific N balances and N deposition were converted to total values beforehand and re-calculated afterwards. The currently processed data addresses different scenarios in NAPSEA (as indicated by the value in the column napsea scenario):
	- o None: baseline for this scenario (napsea\_scenario=2)
	- o DV21: Fertilizer Ordinance not implemented, only anticipated agricultural trends (napsea\_scenario=6), see scenario [3.6.2](#page-16-3)
	- o C\*: DüngEval scenario code, measures beyond Fertilizer Ordinance for scenario 7 (napsea\_scenario=7), see scenario [3.7.1](#page-17-1)
- NL: The modelling units will be intersected with the regions for which scenario data is available to obtain the area-weighted average concentration for agricultural areas in the modelling units [\(Table 5\)](#page-8-1). Since the measures are mainly implemented on sandy and loess soils, the effects of measures in clay soils are assumed to be neglectable.
	- o Reference 2019: current situation as baseline for this scenario
	- o Reference 2027: additional measures not implemented measures for scenario [3.6.2](#page-16-3)
	- o C: most stringent scenario, without voluntary (DWA) measures for this scenario, with voluntary measures for scenari[o 3.7.1](#page-17-1) (alternatively, scenario B)

<span id="page-8-1"></span>*Table 5. Modelled average nitrate concentration in leachate for different regions, sectors as well as the reference year 2019 and scenarios for 2027 (van Boekel et al. 2021) in mg L-1 . The difference between the reference values for 2019 and 2027 reflects the autonomous development of the Dutch agriculture. Scenarios B and C go beyond the planned measures for the scenario 'Reference 2027' and optionally include the effect of voluntary measures of the Dutch Deltaplan Agrarisch Waterbeheer (DAW) program (cf. Gericke and Leujak 2023).*







NOTE: The reference years are identical for all scenarios. The structure of the output files may change to include the area share of the German and Dutch values on the modelling units.

#### <span id="page-9-0"></span>3.2.2. Soil erosion – Soil Health Law

<span id="page-9-3"></span>Spatial coverage: DE x NL x Other x

Nutrient:  $N$  P  $(x)$ 

Scenario: Implementation of the Soil Health Law which sets a maximum total soil erosion of 2 t ha<sup>-1</sup> yr<sup>-1</sup> on arable land for healthy soils (Directorate-General for Environment 2023). The scenario assumes that the reduction of gross soil erosion<sup>[1](#page-9-1)</sup> does not alter the contribution of the erosion processes addressed by the data source (see list below). Additionally, the relative change in the input of particulate P to surface waters equals the change in soil erosion via water and wind erosion as relevant erosion processes.

Data sources: Gridded soil erosion via different processes on arable land (Borrelli et al. 2023), data available from Borrelli et al. (2022). The arable land was obtained from CORINE 2006.

Scenario data: se[e Table 6](#page-9-2)

- Average total soil erosion and coefficient of variation with and without upper threshold [\(Figure 3\)](#page-10-0). In addition to the absolute values (ratio=FALSE), the mean and coefficient of variation was also determined for the ratio scenario/reference (ratio=TRUE).
- Average erosion via the different erosion processes (water erosion, wind erosion, tillage erosion, crop harvesting (SLCH)) as absolute values (ratio=FALSE) or relative to total erosion (ratio=TRUE, [Figure 4\)](#page-10-1)

Data processing: The average total erosion as well as the contribution of the four included erosion processes were obtained for the NUTS 3 regions as soil erosion is hardly relevant for N inputs to surface waters. For the scenarios, the total erosion was capped at the upper threshold of 2 t ha<sup>-1</sup> yr<sup>-1</sup> (for the other thresholds, see scenari[o 3.7.2\)](#page-17-4).

<span id="page-9-2"></span>



<span id="page-9-1"></span><sup>&</sup>lt;sup>1</sup> Without sediment deposition (remobilization) during the transport towards the surface water







<span id="page-10-0"></span>*Figure 3. Total soil erosion on arable land aggregated at NUTS 3 level. The scenario caps the values at 2 t ha-1 .*



<span id="page-10-1"></span>*Figure 4. Contribution of erosion processes to total soil erosion, average values at NUTS 3 level.*





### <span id="page-11-0"></span>3.2.3. Soil erosion – Farm 2 Fork

Spatial coverage: DE x NL x Other x (without Switzerland)

Nutrient:  $N$  P  $(x)$ 

Scenario: 25% organic farming on arable land which reduces the risk of soil erosion due to higher average soil coverage with sod-based crop rotations. The scenario assumes that reduction on soil erosion also applies to the P delivery. As the scenario is based on the C factor of the revised universal soil loss equation, only changes to sheet an[d](#page-11-2) rill erosion by water are considered<sup>2</sup>.

Data sources: Average C factor for Germany and C factor for typical sod-based crop rotation for organic farming provided by Auerswald et al. (2021), current average share of organic farming on arable land (Eurostat 2023)

Data processing: The approach does not allow for spatial variability. If the current C factor of 0.124 represents the current share of organic farming of 5%, and is representative for the whole study areas, increasing the share to 25% would result in C=0.109, i.e. about 12% lower soil erosion if other agricultural changes are neglected (cf. Gericke et al. 2024). For Switzerland, we assume no change in soil erosion as sod-forming crops are also common in conventional farming (Auerswald et al. 2021).

NOTE: The alternative to scenari[o 3.2.2](#page-9-0) assumes an optimal impact as the Farm 2 Fork Strategy refers to agricultural land not arable land.

## <span id="page-11-1"></span>3.3.Atmospheric deposition

Spatial coverage: DE x NL x Other (only EMEP data)

Nutrient: N x P

Scenario: Implementation of the National Emissions Reduction Commitments (NEC) Directive everywhere and the Dutch legislation on Natura2000 areas (NL data), implementation and enforcement of current and planned legislation (EMEP data)

Data sources: Gridded deposition of total nitrogen calculated with PINETI III approach (Schaap et al. 2018) for DE (A. Moravek, pers. comm.), the OPS model (Sauter et al. 2023) for NL (RIVM 2023; Hoogerbrugge et al. 2022), and the EMEP MSC-W model (EMEP MSC-W 2022; Simpson et al. 2012) for the whole study area (EMEP 2022; Denby et al. 2022[\)](#page-11-3)<sup>3</sup>, climate change is not considered

<span id="page-11-4"></span>*Table 7. Structure of output files for the atmospheric deposition with example values. Currently, the modelling units do not cover the Netherlands.*



Scenario data: Average value and coefficient of variation of N deposition in mQM modelling units for the reference years 2015 (DE, NL, EMEP) and 2022 (NL) as well as the scenario years 2030 and 2050 (only EMEP)

<span id="page-11-3"></span><span id="page-11-2"></span><sup>&</sup>lt;sup>2</sup> The empirical universal soil loss equation and its many descendants and adaptations is the most used (family of) models to estimate soil loss rates especially at larger scales (Borrelli et al. 2021) despite inherent flaws (Alewell et al. 2019). Soil loss is the product of 6 factors which reflect the impact of rainfall and runoff (R factor), the soil erodibility (K), the slope angle (S), erosive slope length (L), soil conservation (P), and the soil coverage and crop management (C). The model considers sheet and rill erosion by water but neglects other agents (like wind), processes (gully erosion, landslides), and typically sediment deposition (cf. Borrelli et al. 2022). Typically, the outcome is a long-term average annual soil-loss rate.





[\(Table 7\)](#page-11-4). In addition to the absolute values (ratio=FALSE), the mean and coefficient of variation was also determined for the ratio to the common year 2015 (ratio=TRUE).



<span id="page-12-0"></span>*Figure 5. Distribution of average atmospheric N deposition for mQM modelling units according to national data. The data provider is contacted to confirm the pattern in NL.*



*Figure 6. Distribution of average atmospheric N deposition for mQM modelling units according to EMEP data.*

The available data did not fully cover the modelling units. To allow to calculate area-weighted means for transboundary modelling units, the national average value is provided with its spatial extent. The scenario effect in NL is affected by the inconsistent meteorology of the reference years and the scenario years (2005–2014) which explains the increase between 2022 and 2025 [\(Figure 5\)](#page-12-0).

Data processing: The EMEP data is provided as NetCDF files from which we selected the layers WDEP\_OXN (wet deposition of oxidized nitrogen), WDEP\_RDN (wet deposition of reduced nitrogen), DDEP\_RDN\_m2Grid





(dry deposition of reduced nitrogen per m2 grid), and DDEP\_OXN\_m2Grid (dry deposition of oxidized nitrogen per m2 grid). The total N deposition was calculated as the sum of these layers. From the gridded data, we calculated the average total N deposition and the coefficient of variation for the mQM units and NUTS 3 regions. The units were harmonized to kg ha<sup>-1</sup>.

## <span id="page-13-0"></span>3.4.Nature-based solutions

The scenarios for nutrient retention in floodplains and riparian buffers are work in progress and final scenario data are not yet available. So, new insights might be included in a later stage. The sub-scenarios complement each other. While scenario [3.4.1](#page-13-1) affects the modelled in-stream retention, scenario [3.4.2](#page-13-2) addresses the nutrient input into the surface waters.

#### 3.4.1.Floodplains

<span id="page-13-1"></span>Spatial coverage: DE x NL Other

Nutrient: N x P x

Scenario: Implementation of the EU Nature Restoration Law with 20% more active floodplains in Germany

Data source: Updated dataset after Schulz-Zunkel et al. (2012)

NOTE: The existing database and approach are currently updated (M. Scholz, pers. comm.). For the morphological floodplain along 79 German rivers, a proxy-based approach for the retention potential due to denitrification (N) and sedimentation (P) in the active floodplain was developed and recently improved for N (Kaden et al. 2023). The dataset comprises floodplains with an upstream area above 1000 km<sup>2</sup> (BMU and BfN 2021).

The modelled nutrient retention in tons per year under current conditions and scenario conditions are the sum of all assigned segments (cf[. Figure 7\)](#page-13-3). The results are constant over time and will be provided for floodplain segments which are assigned to the modelling units.



b) Stickstoffretention (Denitrifkation) [in tausend Tonnen]

<span id="page-13-3"></span>*Figure 7. Overall N retention (b) and P retention (c) for reference conditions (Ist-Zustand) and the scenario based on the German National Biodiversity Strategy (Biologische Vielfalt 2010) with 10% more active floodplains where arable land changes to grassland, wetland, and forest along 79 German rivers (source: Schulz-Zunkel et al. 2012). The calculations for the reference conditions and the assumptions for the scenario are currently updated.*

### 3.4.2.Riparian buffers

<span id="page-13-2"></span>Spatial coverage: DE x NL x Other x

Nutrient:  $N(x)$   $P(x)$ 

Scenario: Implementation of §38a of the German Federal Water Act (Wasserhaushaltsgesetz), permanent plant cover of 5 m width on arable land with an average slope of at least 5% within a distance of 20 m from surface waters

Data sources: Land use and land cover (LULC) in riparian zones (EEA 2021), water bodies in EU member states for the reporting within the Water Framework Directive (EEA 2023b), missing water bodies for Luxembourg from a previous version (EEA 2020), digital elevation model (DEM) of 25m-resolution (European Commission – DG ENTR 2012; Eurostat n.d.)





Scenario data: Total area of the arable land adjacent to surface water to be turned into riparian (column 'converted arable land' i[n Table 8\)](#page-14-0) as well as the area of unchanged arable land and other LULC (column 'nonarable land').

The slope threshold results in less new riparian buffers in NL compared to other countries [\(Figure 8\)](#page-14-1) which coincides with the reported low efficiency of narrow grass buffers under Dutch conditions (Noij, Heinen, and Groenendijk 2012).

<span id="page-14-0"></span>*Table 8. Structure of output files for the riparian buffers with example values. Currently, the modelling units do not cover the Netherlands. The area refers to the grid resolution and does not consider the assumed buffer width.*





<span id="page-14-1"></span>*Figure 8. Distribution of share of newly established riparian zones in NUTS 3 regions. The slope threshold results in low values in the Netherlands. The share needs to be adjusted by the average efficiency of riparian buffers to estimate the relative change nutrient retention compared to the current conditions.*

Data processing: The arable land in the riparian zone next to surface waters which is expected to be converted to riparian buffers was derived with a simplified grid-based approach:

- Calculate slope in degree from the DEM
- Extract arable land (value=2) and water areas (value=8) from riparian LULC data (polygon)
- Clip water bodies (polygon, line) along these surface waters and combine the two polygon datasets
- Convert vector data to grid data using the slope grid as spatial reference
- Mark arable land on steep terrain as 0 and on flat terrain as 1
- Expand surface water by one grid cell and mark the arable land there [\(Figure 9\)](#page-15-0)

The slope grid was not resampled as the DEM resolution is close to the required distance of 20 m for the slope calculation. Slopes were considered as 'steep' where the slope angle was above 5%. From the gridded outcome,





we derived the areas of flat and steep arable land as well as other LULC adjacent to surface water for the mQM units and NUTS 3 regions.



<span id="page-15-0"></span>*Figure 9. Output of the raster-based approach (detail) to detect arable land adjacent to surface water which are to be converted to riparian buffers (green) and left unchanged (yellow). For the scenario, we combine the share of the new riparian buffers to the riparian area (grey) with average efficiencies for N and P from the literature.*



<span id="page-15-1"></span>*Figure 10. Median (black) and average (red) efficiency of riparian buffers to retain dissolved (DN, DP) and total (TN, TP) N and P for buffers widths between 4 and 6 m (data: Gericke et al. 2020). The sample sizes are unequal. The 97 literature values are dominated by plot studies.*





NOTE: The approach neglects various site conditions which influence the efficiency of riparian buffers to retain nutrients (e.g. Gericke et al. 2020) as well as the spatial variability of nutrient inputs within the modelling units. It therefore differs from estimations based on e.g. soil type and the presence of drainage (van Boekel et al. 2021).

To estimate the change of nutrient inputs relative to the calibrated models, the value in the column 'converted arable land' needs to be multiplied by a representative average retention factor of 50% for N and P, assuming a buffer width of 5 m [\(Figure 10\)](#page-15-1).

### 3.5.All measures

<span id="page-16-0"></span>The combination of the scenario data above. Exclude the atmospheric deposition in scenario 3.2 to avoid double counting with scenario 3.3.

### <span id="page-16-1"></span>3.6.No measures

<span id="page-16-2"></span>The use of scenario data other than scenario 3.6.1 for this scenario and scenarios [3.1](#page-5-1)[–3.5](#page-16-0) will be discussed.

3.6.1.Climate change

Spatial coverage: DE x NL x Other x

Nutrient:  $N(x)$   $P(x)$ 

Scenario: Hydrological input for ensemble of climate scenarios for the Representative Concentration Pathway (RCP) 6.0 emission scenario calculated with the mHM model (Samaniego et al. 2018). The selected scenario is the medium scenario of the available scenarios for RCP 2.6, RCP 6.0, and RCP 8.5.

Scenario data: Readily available at UFZ, no data processed

<span id="page-16-3"></span>3.6.2. N balances

Spatial coverage: DE x NL Other

Nutrient: N x P

Scenario: Anticipated changes in the agriculture without policy implementation

Data sources: Same as scenari[o 3.2.1](#page-7-2)

<span id="page-16-4"></span>Scenario data: Integrated in data for scenari[o 3.2.1](#page-7-2) (scenario = "DV21", napsea\_scenario = 6)

3.6.3.Erosion – Rainfall erosivity

Spatial coverage: DE x NL x Other x

Nutrient:  $N$  P  $(x)$ 

Scenario: Changes in rainfall erosivity due to changes in rainfall amount and intensity. The scenario is based on the R factor of the revised universal soil loss equation. It is thus limited to sheet and rill erosion by water. The scenario assumes that the relative change in P delivery to surface waters equals the change in soil erosion.

Data sources:

- Gaussian Process Regression ("GPR" in the output file) applied to predict the R factor of the revised universal soil loss equation predominantly for 2010–2020 (Panagos et al. 2015) and around 2041–2060 using the HadGEM global circulation model downscaled with WordClim for RCP 4.5 and the WorldClim climatic datasets as covariates (Panagos et al. 2017), both available from the European Soil Data Centre (Panagos et al. 2022)
- Convection-permitting simulations based on the regional climate model COSMO-CLM for RCP 8.5 emission scenario for 2001–2019 and 2031–2060 (Uber et al. 2024)

Scenario data: Average value and coefficient of variation of the R factor for different time periods [\(Table 9\)](#page-17-5). In addition to the absolute values (ratio=FALSE), the mean and coefficient of variation was also determined for the ratio scenario/reference (ratio=TRUE).





<span id="page-17-5"></span>*Table 9. Structure of output files for the erosivity with example values.*



Data processing: The average R factors were obtained from gridded data for the mQM units and NUTS 3 regions. For the convection-permitting simulation, we created the grids from the published data points by inversedistance weighting using a radius of 10 km.

## <span id="page-17-0"></span>3.7.Adapted

The details of this scenario will be discussed based on the outcomes of scenario[s 3.5](#page-16-0) and [3.6.](#page-16-1) However, exemplary datasets were already created as templates for the processing of scenario data.

#### <span id="page-17-1"></span>3.7.1. N balances

Spatial coverage: DE x NL x Other

Nutrient: N x P

Scenario: Measures beyond scenario [3.2.1](#page-7-2)

Data sources: Same as scenari[o 3.2.1](#page-7-2)

Scenario data: Integrated in data for scenari[o 3.2.1](#page-7-2) (e.g. scenario code starts with "C" in Germany, napsea\_scenario = 7). The DüngEval results with the most notable Germany-wide effects were pre-selected:

- Scenario = C08: amount of N fertilizer reduced to 80 % of plant demand (currently required in hotspot  $($ "red") areas)
- Scenario = C09: site-specific fertilization
- Scenario = C14: conversion to and extensification of permanent grassland within water protection areas

Depending on the required ambition, the Dutch scenarios B (less stringent) and C (more stringent) can be chosen.

<span id="page-17-2"></span>3.7.2. Goal of Nitrates Directive achieved

<span id="page-17-4"></span>Spatial coverage: DE x NL x Other x

Nutrient: N x P

Scenario: In addition to scenario[s 3.2.1](#page-7-0) an[d 3.7.1,](#page-17-1) we assume in the mQM model that the N concentration in the groundwater (seepage water) is 50 mg  $L^{-1}$  N.

Scenario data: No model input data needed.

<span id="page-17-3"></span>3.7.3. Soil erosion

Spatial coverage: DE x NL Other

Nutrient:  $N$  P  $(x)$ 

Scenario: Upper threshold of total soil erosion of the Soil Health Law reduced to 1.5 t ha<sup>-1</sup> yr<sup>-1</sup> which is slightly above upper limit of soil formation rate in Europe (Verheijen et al. 2009) and 1.0 t ha<sup>-1</sup> yr<sup>-1</sup> as recommended by the same authors when considering the impact of soil erosion on water quality

Data sources: Same as scenari[o 3.2.2](#page-9-3)





Scenario data: Integrated in data for scenari[o 3.2.2,](#page-9-3) the scenario column represents the upper threshold, napsea\_scenario = 7

### <span id="page-18-0"></span>3.7.4. Change of land use and land cover (LULC)

Spatial coverage: DE x NL x Other x

#### Nutrient:  $N(x)$   $P(x)$

Scenario: Change of LULC related to various RCP and Shared Socioeconomic Pathways as used for the 6th phase of Coupled Model Intercomparison Project (CMIP6) (Hoffmann et al. 2023) as well as under various Nature Futures Framework scenarios to assess the impact of sustainability targets (Dou et al. 2023).

Data sources:

- Gridded LUCAS LULC change dataset with simulated annual maps 1950–2100 (Hoffmann et al. 2023), data available from Hoffmann et al. (2022)
- Gridded LULC data with land-use intensity (low-medium-high) for scenarios in 2050 using an adapted map of Dou et al. (2021) as starting point representing the year 2015 (Dou et al. 2023), data available from Verburg (2023)

Scenario data: Two similar data tables (Table[s 10](#page-18-1)[–11\)](#page-18-2) with values for the respective reference and scenario years

<span id="page-18-1"></span>*Table 10. Structure of output files for the LULC data provided by Hoffmann et al. (2022).*



<span id="page-18-2"></span>*Table 11. Structure of output files for the LULC data provided by Verburg (2023). The grid values (LULC class) were adjusted to match the range of grid values.*







#### Processing:

- First dataset: Given the climate data for scenario [3.6.1,](#page-16-2) we selected the results for RCP 6.0. The RCP was combined with the Shared Socioeconomic Pathway "Inequality (A Road Divided)". For each year between 2016 and 2055 and the NUTS / mQM units, we obtained the average area share from the grid data as well as the coefficient of variation.
- Second dataset: The value range (0–20) in the description of the grid values in Verburg (2023) did not match the actual range of grid values (0–20, 0–21, 1–21). After visual inspection, the grid values were harmonized, and the description extended by the missing entry for water (class value  $= 0$ ). For each area id, we derived the spatial extent of all (adjusted) land-use classes. All published datasets were processed and indicated by the SSP and scenario columns in the output file. The Nature Futures Framework scenarios nac, nfn, and nfs are based on the SSP1 scenario and focus on Nature's
	- $\circ$  non-material contributions to society, priority are cultural services (scenario = nac)
	- $\circ$  benefits for society, priority is mitigation of climate change (scenario = nfs)
	- $\circ$  intrinsic value, priority is distribution and protection of vertebrates (scenario = nfn)

### <span id="page-19-0"></span>3.8.Discussion

The outcome of text files cannot be directly used as input for the NAPSEA models mQM and C<sup>n</sup>ANDY. Firstly, the available data typically represent different single years in the future while we agreed during the workshop with German stakeholders on April 25, 2024 to model the scenarios until 2050. This issue requires for the mQM model assumptions on how the scenario is achieved, e.g. steadily or stepwise. Secondly, the datasets also differ from the data used to calibrate the models. Therefore, the original model input for the reference year of the scenario should be adjusted by the (relative or absolute) change between the reference and future years in the scenario data. As N balances in kg ha<sup>-1</sup> depend on the extent of agricultural land in the maps used for the model input and the scenario data, absolute values (in kg) should be derived from the area-specific values.

The values at NUTS 3 level are intended as provisional input for the C<sup>n</sup>ANDY. Depending on the model requirements regarding the data resolution, the current data could be further aggregated to obtain an average change of P export from the land-use classes or replaced by the raster data itself. Unlike the first requirement, the second one would require changes to the data processing. As mQM currently does not cover NL, all data files with N data do currently not contain data (either as rows or columns) with Dutch data. However, these issues can be rapidly addressed with the established scripts.

Technically, the finalization of the scenario data requires only a few basic steps. However, we have to discuss how to fill gaps and how to harmonize different scenario inputs once the mQM model covers the Netherlands. This is pivotal for the agricultural N input given its dominance over other sources (cf. Gericke and Leujak 2023).

The available model data and scenarios are inherently uncertain. For instance, Häußermann et al. (2019) quantifies uncertainties in German N balances which are supposedly valid for the values obtained for scenarios [3.2.1,](#page-7-0) [3.6.2,](#page-16-3) an[d 3.7.1.](#page-17-1) The UWWTD demands from EU member states that 20% of the agglomerations meet by end of 2033 the targets of scenari[o 3.1,](#page-5-1) 40% by 2036, 60% by 2039, and 100% by 2045. Even if we assume the implementation of the revised UWWTD, it remains unclear which UWWTP will achieve the envisioned targets in these years. To which degree such uncertainties should be explicitly considered in the modelling needs to be discussed during the finalization of the scenario data. To support such analyses, we included the coefficients of variation of some of the average values.





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## <span id="page-23-0"></span>ANNEX

<span id="page-23-1"></span>*Table 12. Preliminary set of scenarios for the Hunze catchment (J. Rozemeijer, pers. comm., slightly modified)*

